Geostatistical Tools for Assessing Sampling Designs Applied to a Portuguese Bottom Trawl Survey Field Experience

 $Ernesto \ Jardim^1 < ernesto@ipimar.pt >$

Paulo J. Ribeiro Jr² <paulojus@ufpr.br>

October 23, 2007

¹Instituto Nacional de Investigação Agrária e das Pescas, Av. Brasilia, 1449-006, Lisboa, Portugal ²Laboratório de Estatística e Geoinformação, Universidade Federal do Paraná, C.P. 19.081 CEP: 81.531-990, Curitiba, Paraná, Brasil

1

Abstract

This paper presents a bottom trawl survey (BTS) field experience carried out off the Portuguese Continental shelf to test two sampling designs proposals previously analysed by simulation (Jardim and Ribeiro Jr, 2007) which implement an hybrid random-systematic and a systematic sampling strategy. We used a common base regular grid covering the survey area and overlap it with the existent random design to build the hybrid design while the systematic design adds a set of regular locations at smaller distances creating four denser sampling areas. We use hake (Merluccius merluccius) yield and model-based geostatistics (Diggle and Ribeiro Jr, 2007) to compute tools like: mean abundance, μ , • and the 95% percentile, p_{95} , that summarise the areal behaviour; coverage of the prediction confidence interval, ξ , to assess the adequacy of the model; and a modified generalised cross validation index, 10 ε , to evaluate prediction precision. The hybrid design showed a lower coefficient of variation for μ 11 (11.89% against 13.25%; a slightly higher coefficient of variation for p_{95} (11.31% against 11.09%; 12 similar ξ (0.94); and lower ε (16.32 against 18.82). We conclude that: the hybrid design performs 13 better and our procedure to build it can be used to adjust BTS designs to modern geostatistical 14 techniques; and the statistics used constitute valuable tools to assess BTS performance. 15

16 Key-words: model-based; geostatistics; hake; sampling design; bottom trawl survey.

17 1 Introduction

Designs for Bottom trawl survey (BTS) rely on previous knowledge of the target species regarding spatial 18 distribution and population structure combined with statistical analysis of preliminary data (e.g. Ault et al., 1999; Hata and Berkson, 2004) or simulation procedures (e.g. Schnute and Haigh, 2003; Anon., 20 2005b). These results are confronted with operational constraints such as trawlable grounds and vessel 21 availability, among others, to define the definitive BTS sampling design. The survey design is typically 22 reviewed from time to time to adjust the stratification (e.g. Smith and Gavaris, 1993; Folmer and Pen-23 nington, 2000), tow duration (e.g. Cerviño and Saborido-Rey, 2006; Wieland and Storr-Paulsen, 2006), 24 technical issues such as gear changes (e.g. Zimmermann et al., 2003; Cooper et al., 2004) and other factors 25 which may change over the years.

The Portuguese BTS started in June 1979, covering the continental shelf and following a stratified random 27 design. In 1989 the stratification was defined by 12 sectors along the coast subdivided into 4 depth ranges: 28 20-100m, 101-200m, 201-500m and 501-750 m, with a total of 48 strata. Due to constraints in the vessel 29 time available the sample size was set to 97 locations evenly allocated to each stratum. The coordinates 30 of the sampling locations were selected randomly, albeit constrained by the historical records of clear 31 tow positions and other information about the sea floor, thus avoiding places where trawling was not 32 possible. During this period haul duration was set to one hour but recent experiments proved that half 33 hour hauls provide the same information about length distributions (Cardador, pers.comm.). In light of 34 this findings haul duration was reduced to half hour and an additional set of hauls were available which 35 motivated a revision of the sampling design. The revision was splitted into a preliminary phase using simulations and geostatistical analysis (Jardim and Ribeiro Jr, 2007) and a second phase during which 37 a field test was executed to provide real information about the proposed sampling designs. In a third 38 moment the decision will have to be made based on the scientific data provided and the existing financial 39 and administrative constraints. 40

The field experience was carried out during the summer of 2001, with R/V Noruega off the southwest of Portuguese Continental shelf (Fig. 1) using a Norwegian Campbel Trawl 1800/96 (NCT) with a codend of 20mm, mean vertical opening of 4.8m and mean horizontal opening between wings of 15.5m. The survey executed two sampling designs selected from the simulation study reported by Jardim and Ribeiro Jr (2007). The survey area was limited on the south by the cape of S.Vicente (37.00° north), on the north by Setubal's Canyon (38.30° north), on the east by the 20m depth isoline and on the west by the 500m isoline. The survey area had approximately 4300km² and the maximum distance within the area was approximately 150km. The data collected on both designs and considered here consists of date/time, geographical location and hake (*Merluccius merluccius*) catch in weight (kg). Geographical coordinates were transformed into UTM units and hake abundance was computed in kg/km and assigned to the haul starting coordinates. The area swept was computed using the haul start and ending positions to correct haul speed variations.

Our analysis adopts model-based geostatistical method (Diggle et al., 1998; Diggle and Ribeiro Jr, 2007) 53 to explicitly take into account spatial patterns of abundance and provide a flexible modelling framework. The designs are accessed by a set of statistics to provide information about different aspects of the data, relevant for modelling fish abundance. In a global perspective, referring to the entire study region, we use mean abundance and the 95% percentile to summarise the areal behaviour of abundance, commonly used 57 for studying time trends and building abundance indices for stock assessment. In a local perspective, 58 referring to particular locations within the study area, we use the observed values to assess the adequacy 59 of the model, computing the coverage of the prediction confidence interval, and the prediction precision, 60 computing a modified generalised cross validation index. Note that the assessment of the model adequacy 61 and the prediction precision are extremely valuable statistics, once that kriging is in fact a linear predictor 62 and the maps produced with it will be used to estimate the spatial distribution of abundance and the 63 abundance index mentioned before. With relation to the analysis reported here we rely on our experience with bottom trawls surveys (Anon., 2002, 2003, 2004, 2005a, 2006; Sousa et al., 2005; Mendes et al., 2007; Sousa et al., 2007) to provide contextual information which supports the adoption of a particular class of models, and avoid as much as possible model mis-specification. 67

The work described on this paper aims at: (i) reporting a BTS field experience to test sampling designs, and (ii) describe geostatistical tools to assess the performance of sampling designs. Although the results obtained are constraint by the characteristics of the area and the species analysed, we believe the methodology defined by our approach can be applied to other areas and species, providing an important source of information when revising sampling design.

$_{73}$ 2 Methods

This section describes the sampling designs to be tested and how they were built. It also describes
the geostatistical modelling framework and the adjustments considered to cope with the small dataset
available, a common characteristics of BTS due to its high price. At last we describe the technical details
of the performance statistics chosen.

78 2.1 Sampling designs

Several authors discussed the advantages of systematic designs over random designs to sample spatial 79 correlated variables like fish abundance (Cochran, 1960; Ripley, 1981; Thompson, 1992; Cressie, 1993; 80 Chiles and Delfiner, 1999; Kimura and Somerton, 2006; Diggle and Ribeiro Jr, 2007). Nevertheless, in 81 the case of spatial correlated variables there are two conflicting objectives that can not be combined 82 in a single criteria, estimation of the covariance function parameters and prediction (Muller, 2001). In 83 the first situation it is important to have locations at short distances to inspect the behaviour of the 84 correlation function close to the origin, and locations at distances close to the limit of spatial correlation 85 to estimate the correlation range (Muller, 2001). In the second situation the best predictions will result 86 from the design with higher covariance with the locations to be predicted (Thompson, 1992). In the case 87 of predicting fish abundance it is common to require a complete map of the study area and the best 88 choice will be a design that covers the area evenly. However, when the covariance function is unknown, a common characteristic of fish abundance analysis, it must be estimated from the data before predicting and both objectives must be combined. Several authors propose designs that mix a set of locations covering the area with additional locations at short distances (Muller, 2001; Diggle and Lophaven, 2006; 92 Zhu and Stein, 2006) to balance between both objectives. Such designs were not considered for bottom 93 trawl surveys until now, although fish abundance characteristics fit well in the assumptions of these 94 proposals. Our sampling designs were built mixing a set of operational constraints with the geostatistical 95 principles elaborated above and the need to keep the continuity of the survey history. In particular, 96 the two designs tested were built to distinguish between an hybrid random-systematic sampling strategy 97 and a systematic strategy. Both designs were built from a common basis, a regular grid covering the 98 survey area. The hybrid design overlaps this regular grid with the existent random design keeping some ٩q continuity with the survey historical records (top-left plot in Fig. 2). The systematic design includes a 100 set of locations positioned regularly at smaller distances creating 4 denser sampling areas (bottom-left 101 plot in Fig. 2). 102

¹⁰³ 2.2 Geostatistical model

Geostatistical observations consist of pairs (x, y) with elements $(x_i, y_i) : i = 1, ..., n$, where x_i denotes the coordinates of each of the *n* spatial locations within a study region $A \subset \mathbb{R}^2$ and y_i the measurement of the corresponding observable study variable. We adopted the Box-Cox transformed Gaussian model with transformation parameter λ as presented in Christensen et al. (2001). Denoting by z_i the transformed values, such that $g_{\lambda}(y_i) = z_i$, the model for the vector of variables Z observed at locations x can be written as a linear model $Z(x) = S(x) + \epsilon$, where S is a stationary Gaussian stochastic process, with

 $E[S(x)] = \mu$, $Var[S(x)] = \sigma^2$ and an isotropic correlation function $\rho(h) = Corr[S(x), S(x')]$, where 110 $h = \|x - x'\|$ is the Euclidean distance between locations x and x'. The terms ϵ are assumed to be 111 mutually independent and identically distributed, $\epsilon \sim \text{Gau}(0, \tau^2)$. For the correlation function $\rho(h)$ we 112 adopt the exponential function with algebraic form $\rho(h) = \exp\{-h/\phi\}$ where ϕ is the range parameter 113 such that $\rho(h) \simeq 0.05$ when $h = 3\phi$. Following usual geostatistical terminology (Isaaks and Srivastava, 114 1989) we call $\sigma_T^2 = \tau^2 + \sigma^2$ the total sill, σ^2 the partial sill, τ^2 the nugget effect and 3ϕ the practical range. 115 Geometric anisotropy (Isaaks and Srivastava, 1989; Cressie, 1993) is considered an extension of this model 116 with extra parameter $\psi = \{\psi_A, \psi_R\}$ where ψ_A is the anisotropic angle and ψ_R is the anisotropic ratio. 117

Hereafter we use $[\cdot]$ to denote the distribution of the quantity indicated within brackets. Following the 118 adopted model, $[g_{\lambda}(Y)] \sim \text{MVGau}(\mu \mathbf{1}, \Sigma)$, i.e. [Y] is multivariate trans-Gaussian with expected value μ 119 and covariance matrix Σ parametrised by $\{\sigma^2, \phi, \tau^2\}$. Parameter estimates can be obtained by maximum 120 likelihood (Cressie, 1993; Diggle et al., 1998; Diggle and Ribeiro Jr, 2007) and used for spatial prediction. 121 In its simplest format, spatial prediction given by the kriging predictor consists of obtaining expected 122 values and associated variances at unsampled locations. More generally, the *predictive distribution* of 123 quantities of interest can be obtained analytically, if possible, or by sampling from this distribution. Con-124 sider a prediction target $T(x_0) = g_{\lambda}^{-1}(S(x_0))$, the realised value of the process in the original measurement 125 scale at spatial locations x_0 . Simulations from the conditional distribution $[T(x_0)|Y(x)]$ are obtained by 126 simulating from the multivariate Gaussian $[S(x_0)|Y(x)]$ and back transforming the simulated values to 127 the original scale of measurement (Chiles and Delfiner, 1999; Diggle and Ribeiro Jr, 2007). These simu-128 lations are called *conditional simulations* referring to the fact they are obtained from the distribution of 129 the quantity of interest conditioned to the observed values Y(x). 130

We split inference in two steps. First the Box-Cox transformation parameter λ and the anisotropy pa-131 rameter ψ_R are investigated by pooling all the observations in a single dataset and computing profile 132 likelihoods (Diggle and Ribeiro Jr, 2007). We consider the north-south coastal orientation of the study 133 region as the direction of greater spatial continuity and fix ψ_A in 0 degrees azimuthal angle. Afterward, 134 having estimated these two parameters we regard their point estimates as constants in the model and 135 proceed by computing, for each design, the maximum likelihood estimates of the remaining model param-136 eters. The reasoning for the two steps procedures is twofold. Pragmatically, this overcome the difficulty 137 to identify all parameters with a small dataset, whereas in terms of modelling assumptions we regard 138 the transformation and anisotropy parameters as part of the model specification, reflecting the nature 139 of the data and contextual information and therefore not to be identified by the designs. Thereafter, we 140 compute kriging predictions on a 2×2 km grid within the study area, x_0 , with a total of 1070 locations, 141 and obtain 1,000 conditional simulations from $[Y(x_0)|Y]$ for each design. 142

143 2.3 Performance statistics

Consider $E[Z(x_i)]$ and $\sigma_z^2(x_i)$ the kriging predictor and its variance on the Gaussian scale at location 144 $x_i \in x_0$ and the transformation parameter $\lambda = 0.5$. Back transformation to the original scale gives $E[Y(x_i)] = (1 + 0.5E[Z(x_i)])^2 + 0.25\sigma_z^2(x_i)$ and the global mean is estimated by averaging the predicted 146 values $\hat{\mu} = m^{-1} \sum_{i=0}^{m} \hat{E}[Y(x_i)]$. The variance of $\hat{\mu}$, denoted by $\hat{\sigma}_{\mu}^2$, is computed by the sum of all terms in 147 the covariance matrix $\Sigma_Y(x_0) = Var[Y(x_0)|Y(x)]$, back transformed by $\Sigma_Y(x_0) = \Sigma_Z(x_0)[8^{-1}\Sigma_Z(x_0) + 12^{-1}\Sigma_Z(x_0)]$ 148 $(1 + 0.5E[Z(x)])^2$, where $\Sigma_Z(x_0)$ is the covariance matrix of $[S(x_0)|Z(x)]$. More generally, inferences 149 on other quantities of interest $T(x_0)$ are obtained from the conditional simulations. Denote by $t_s(x_0)$, 150 $s = 1, \ldots, S = 1,000$ conditional simulations from $[T(x_0)|Y(x)]$. For example, an α -th percentile is 151 estimated by $\hat{p} = S^{-1} \sum_s \hat{p}_s$ where $\hat{p}_s = p_\alpha(t_s(x_0))$, the average of the empirical distribution \hat{p} obtained 152 from the conditional simulations. The variance of \hat{p} is given by $\hat{\sigma}_p^2 = (S-1)^{-1} \sum_s (\hat{p}_s - \hat{p})^2$. 153

The coverage of the prediction confidence interval, ε , and the generalised cross validation index, ξ , were 154 computed using cross-validation statistics (Hastie et al., 2001) combined with conditional simulations as 15! follows. First, create a new data set by leaving one observation out at a location x_i , simulate 1,000 values 156 of the variable at that location, and repeated this procedure visiting all data locations. Subsequently, con-157 sider $y(x_i)$ an observation of the process Y on location x_i , i = 1, ..., n; $y(x_{(i)})$ the observed data set with-158 out the observation $y(x_i)$ and $t_s(x_i)$ a conditional simulation s = 1, ..., S of $[T(x_i)|Y = y(x_{(i)})]$ on loca-159 tion x_i . The predictive confidence interval is given by $CI(x_i) = [p_{2.5}(t_s(x_i)), p_{97.5}(t_s(x_i))]$ and the propor-160 tion of observations lying inside the intervals $\xi = n^{-1} \sum_i (y(x_i) \in CI(x_i))$ provides the *coverage* of the pre-161 diction confidence interval. The cross validation index is given by $\varepsilon = n^{-1} \sum_{i} (S^{-1} \sum_{s} (t_s(x_i) - y(x_i))^2),$ 162 the average of the mean quadratic error on each location estimated using the full set of conditional 163 simulations. 164

165 **3** Results

The two sampling designs and the observations of hake yield are presented in the leftmost panels of Figure 2 where the base regular design is represented by the black triangles. The abundance of hake observed showed that the distribution of yield was spread over the area, presenting lower values in the north and a small number of zeros.

The 95% confidence interval obtained for the Box-Cox transformation parameter was [0.12, 0.55] and we set $\hat{\lambda} = 0.5$, corresponding to a square root transformation. The profiled log-likelihood of the anisotropy ratio showed no evidence of anisotropy. Nevertheless, we carried out analysis using different values of ψ_R to check the sensibility of the results, which proved negligible. Covariance parameters estimates presented higher values for the hybrid design than the corresponding ones given by the systematic design (Table 1). The total variance $\hat{\sigma}_T^2$ was 3.75, with $\hat{\tau}^2 = 0.75$ and $\hat{\sigma}^2 = 3.00$; and $\hat{\phi} = 16.64$. While the systematic design estimates were $\hat{\sigma}_T^2 = 3.20$, with $\hat{\tau}^2 = 0.61$ and $\hat{\sigma}^2 = 2.59$; and $\hat{\phi} = 10.21$. Looking at τ_{REL}^2 and $\sigma^2 \phi^{-1}$, that give information about the variability of the spatial process, both designs showed similar relative nuggets but the hybrid design showed a lower ratio between sill and range, reflecting a higher spatial structure of the stochastic process. Notice that the practical range, 3ϕ , was $\approx 50km$ for hybrid and $\approx 30km$ for the systematic design.

The rightmost panels of Figure 2 show the abundance maps predicted and their variance, for each design. Both predictions are similar and the spatial pattern of variance reflects the influence of the observations, showing lower variability near the observed locations and higher variability in areas where extrapolation was further extended. The hybrid design had higher variance in the centre-east of the study area and lower variance on the north due to a better coverage in this area.

The estimates of μ and p_{95} were similar although the hybrid design presented slightly lower values. The 186 hybrid design showed a lower coefficient of variation for μ , $CV_{\mu} = 11.89\%$ than the systematic design, 187 $CV_{\mu} = 13.25\%$. The p_{95} variance was slightly lower for the systematic design, $CV_{p95} = 11.09\%$, while the 188 hybrid design presented $CV_{p95} = 11.31\%$. The coverage of the prediction confidence intervals was 0.94 189 for both designs. These results reinforce our modelling choices given that if the model was wrong we'd 190 expect ξ to be different from the nominal value of the confidence interval. The generalised cross validation 1 9 1 index presented a lower estimate with the hybrid design, 16.32, than with the systematic design, 18.82, 192 showing an higher prediction precision of the hybrid design. The above mentioned results reflect that 193 the higher spatial structure of the stochastic process estimated for the hybrid design surpassed its higher 194 total variability with relation to the estimation of these performance statistics. 195

¹⁹⁶ 4 Discussion

Assessing sampling designs for BTS raises interesting questions about appropriated methodologies to analyse data and derive statistics of interest, which are particularly relevant considering the multipurpose/multispecies nature of BTS and the small sample sizes.

The adoption of a formal criteria and loss function to find an optimum design seems unrealistic in practice due to the multidimensionality of the data and the conflicting objectives of inference and prediction. Here we follow a pragmatic approach to sampling design and started by choosing a design that joins a regular grid with the old random design, followed by a second design that uses the same regular grid but reallocates the random locations in a regular shape. This way we build designs that implement the two most promising strategies, considering the wide literature that support the use of systematic designs for spatial correlated variables, and test the possibility of keeping the continuity with the historical time series. To compare these proposals we rely on spatial modelling to compute statistics of primary interest and look for consistency among them, exploring several aspects of the same dataset. We advocate that the approach described above will provide valuable information to support the decision process.

The performance statistics were selected to reflect relevant characteristics and different aspects of spatial 210 prediction. The global mean is the most used index of abundance, often estimated by the sample average. 211 We favour the geostatistical estimator presented and its variance as a measure of uncertainty, considering 212 it takes into account the spatial dependency within the area and insights about the spatial process. The 21 3 95th percentile estimated by conditional simulations can be used to identify areas of high abundance, 214 giving information about candidate areas to protect. The coverage of the prediction confidence intervals 215 is a diagnostic tool. A small coverage reflects an underestimation of the variance or the inadequacy of the 21 6 model to explain the available data. The cross validation index combined with conditional simulations, 217 incorporates the prediction precision in the index, which is not taken into account by the traditional 218 cross validation. For example, if a location has the same predicted value by different designs but with 219 different prediction variances, our approach would distinguish both situations, differently from the usual 220 cross validation index. 221

Our results showed that the hybrid design performed better in all cases except for σ_p^2 . A clear parallel 222 can be established with the *lattice plus closed pairs* designs of Diggle and Lophaven (2006), the EK-223 optimal designs of Zimmerman (2006) or the D_{EA} designs of Zhu and Stein (2006). All of these cover the 224 study area and include a set of positions at small distance, albeit following different constructions, these 225 designs performed better than their random or systematic competitors. Common to all these studies and 226 our work, is the fact that the analysis were carried out in situations where the model parameters were 227 considered unknown and needed to be estimated from the data, which made it clear that both parameter 228 estimation and prediction are important for the precision of the prediction target. 229

²³⁰ Concluding, we consider that our results give indications that keeping the old random design and add
²³¹ a regular grid to build a new design can be a good and pragmatic solution to adjust BTS designs to
²³² modern model-based geostatistics techniques. Secondly, the performance statistics described above seem
²³³ to capture the most important features of the data with relation to abundance estimation, constituting
²³⁴ good measures to assess BTS performance.

5 Acknowledgements

The authors would like to thank the scientific teams involved in the Portuguese Bottom Trawl Surveys, in particular the coordinator Fátima Cardador. This work was carried out within scope of the IPIMAR's

project NeoMAv (QCA-3/MARE-FEDER, http://ipimar-iniap.ipimar.pt/neomav). 238

References 239

- Anon. 2002. Report of the International Bottom Trawl Survey Working Group. Tech. rep., International 24 0 Council for the Exploitation of the Sea (ICES). ICES CM 2002/D:03. 24
- Anon. 2003. Report of the International Bottom Trawl Survey Working Group. Tech. rep., International 242
- Council for the Exploitation of the Sea (ICES). ICES CM 2003/D:05.
- Anon. 2004. Report of the International Bottom Trawl Survey Working Group. Tech. rep., International 244 Council for the Exploitation of the Sea (ICES). ICES CM 2004/D:05.
- Anon. 2005a. Report of the International Bottom Trawl Survey Working Group. Tech. rep., Interna-246 tional Council for the Exploitation of the Sea (ICES). ICES CM 2005/D:05. 247
- Anon. 2005b. Report of the Workshop on Survey Designs and Data Analysis. Tech. rep., International 248 Council for the Exploitation of the Sea (ICES). 24 9
- Anon. 2006. Report of the International Bottom Trawl Survey Working Group. Tech. rep., International 250 Council for the Exploitation of the Sea (ICES). ICES CM 2006/D:05. 251
- Ault, J., Diaz, G., Smith, S., Luo, J. and Serafy, J. 1999. An Efficient Sampling Survey Design to 252 Estimate Pink Shrimp Population Abundance in Biscayne Bay, Florida. North American Journal of 253
- Fisheries Management, , 19: pp. 696–712. 254
- Cerviño, S. and Saborido-Rey, F. 2006. Using bootstrap to investigadte the effects of varying tow lengths 255 And catch sampling schemes in fish survey. Fisheries Research, , 79: pp. 294-302. 256
- Chiles, J.-P. and Delfiner, P. 1999. Geostatistics: Modeling Spatial Uncertainty. Wiley, New York. 257
- Christensen, O., Diggle, P. and Ribeiro Jr, P. 2001. Analysing positive-valued spatial data: the trans-258
- formed Gaussian model. In: P. Monestiez, D. Allard and Froidevaux (eds.), GeoENV III Geostatistics 259
- for Environmental Applications, vol. 11 of Quantitative Geology and Geostatistics. Kluwer, pp. 287-260 298.
- Cochran, W. 1960. Sampling Techniques. Statistics, John Wiley and Sons, New York. 262
- Cooper, A., Rosenberg, A., Stefánson, B. and Mangel, M. 2004. Examining the importance of consistency 263
- in multi-vessel trawl survey design based on the U.S. west coast groundfish bottom trawl survey. 264
- Fisheries Research, 70: pp. 239–250. 265

261

- ²⁶⁶ Cressie, N. 1993. Statistics for spatial data Revised Edition. John Wiley and Sons, New York.
- ²⁶⁷ Diggle, P. and Ribeiro Jr, P. 2007. Model-based Geostatistics. Springer, New York, 228 pp.
- Diggle, P., Tawn, J. and Moyeed, R. 1998. Model-based geostatistics (with discussion). Appl. Statist.,
 47: pp. 299-350.
- Diggle, P. J. and Lophaven, S. 2006. Bayesian geostatistical design. Scandinavian Journal of Statistics,
 33: pp. 55-64.
- Folmer, O. and Pennington, M. 2000. A statistical evaluation of the design and precision of the shrimp trawl survey off West Greenland. Fisheries Research, , 49: pp. 165–178.
- Hastie, T., Tibshirani, R. and Friedman, J. 2001. The Elements of Statistical Learning. Data Mining,
 Inference, and Prediction. Springer Series in Statistics, Springer, New York, 533 pp.
- Hata, D. and Berkson, J. 2004. Factors Affecting Horseshoe Crab Limulus polyphemus Trawl Survey
 Design. Transactions of the American Fisheries Society, 133: pp. 292–299.
- Isaaks, E. and Srivastava, M. 1989. An Introduction to Applied Geostatistics. Oxford University Press,
 New York.
- Jardim, E. and Ribeiro Jr, P. 2007. Geostatistical Assessment of Sampling Designs for Portuguese
 Bottom Trawl Surveys. Fisheries Research, 85: pp. 239–247.
- Kimura, D. K. and Somerton, D. A. 2006. Review of Statistical Aspects of Survey Sampling for Marine
 Fisheries. Reviews in Fisheries Science, 14, 3: pp. 245–283.
- Mendes, J. M., Turkman, K. F. and Jardim, E. 2007. A Bayesian hierarchical model for over-dispersed
- count data: a case study for abundance of hake recruits. Environmetrics, 18, 1: pp. 27-53. URL
 http://dx.doi.org/10.1002/env.800.
- ²⁸⁷ Muller, W. 2001. Collecting Spatial Data Optimum Design of Experiments for Random Fields. ²⁸⁸ Contributions to statistics, Physica-Verlag, Heidelberg, 2nd ed.
- Ripley, B. 1981. Spatial Statistics. Probability and Statistics, Wiley, New Jersey, 252 pp.
- Schnute, J. and Haigh, R. 2003. A simulation model for designing groungfish trawl surveys. Canadian
 Journal of Fisheries and Aquatic Science, , 60: pp. 640–656.
- ²⁰² Smith, S. and Gavaris, S. 1993. Improving the Precision of Abundance Estimates of Eastern Scotian
- Shelf Atlantic Cod from Bottom Trawl Surveys. North American Journal of Fisheries Mangement, ,
 13: pp. 35–47.

- Sousa, P., Azevedo, M. and Gomes, M. 2005. Demersal assemblages off Portugal: Mapping, seasonal, 295 and temporal patterns. Fisheries Research, 75: pp. 120–137. 296
- Sousa, P., Lemos, R., Gomes, M. and Azevedo, M. 2007. Analysis of horse mackerel, blue whiting, and 297
- hake catch data from Portuguese surveys (1989-1999) using an integrated GLM approach. Aquatic
- Living Resources, 20: pp. 105–116. 299

298

- Thompson, S. K. 1992. Sampling. Probability and Methematical Statistics, Wiley, New York, 343 pp. 300
- Wieland, K. and Storr-Paulsen, M. 2006. Effect of tow duration on catch rate and size composition 301
- of Northern shrimp (Pandallus borealis) and Greenland halibut (Reinhardtius hippoglossoides) in the 302
- West Greeland Bottom Trawl Survey. Fisheries Research, , 78: pp. 276-285. 303
- Zhu, Z. and Stein, M. 2006. Spatial Sampling Design for Prediction With Estimated Parameters. Journal 304 of Agricultural, Biological, and Environmental Statistics, 11: pp. 24-44. 305
- Zimmerman, D. 2006. Optimal network design for spatial prediction, covariance parameter estimation, 306 and empirical prediction. Environmetrics, 17: pp. 635–652. 307
- Zimmermann, M., Wilkins, M., Weinberg, K., Lauth, R. and Shaw, F. 2003. Influence of improved 308 performance monitoring on the consistency of a bottom trawl survey. ICES Journal of Marine Science, 309 , 60: pp. 818-826. 31 0

Table 1: Estimates of model parameters and performance statistics by design. Model parameters are: τ^2 , the short distance variance or nugget effect; σ^2 the variance of the spatial process; σ_T^2 the total variance; ϕ the correlation range parameter; and the transformation parameters λ , the Box-Cox parameter and the anisotropy parameters $\{\psi_A, \psi_R\}$. The relative nugget, τ_{REL}^2 , and the ratio between relative sill and range $\sigma^2 \phi^{-1}$, were computed to give more insights about the spatial process. Performance statistics are: $\bar{\mu}$ and $\sigma_{\bar{\mu}}^2$, the mean and variance of the global abundance; \hat{p}_{95} and $\hat{\sigma}_p^2$, the mean and variance of the 95th percentile of the global abundance; ε , the generalised cross validation index and ξ , the coverage of the prediction confidence interval with nominal level of 0.95.

	hybrid	systematic
model parameters		
$ au^2$	0.75	0.61
σ^2	3.00	2.59
σ_T^2	3.75	3.20
ϕ	16.64	10.21
$ au_{REL}^2$	0.20	0.19
$\sigma^2 \phi^{-1}$	0.18	0.25
ψ_A	0.00	0.00
ψ_R	1.00	1.00
λ	0.50	0.50
performance statistics		
$\hat{\mu}$	4.07	4.20
$\hat{\sigma}_{\mu}^2$	0.23	0.31
cv	11.89	13.25
\hat{p}_{95}	11.01	10.78
$\hat{\sigma}_n^2$	1.55	1.43
cv	11.31	11.09
ξ	0.94	0.94
ε	16.32	18.82



Figure 1: Survey area on the southwest of the Portuguese Continental shelf between 20m and 500m.

Figure 2: Study area on the Portuguese southwest coast. The top panels show information about the hybrid random-systematic design and the bottom panels about the systematic design. The leftmost plots show the sampling designs locations, the black triangles represent the regular grid common to both designs, and the open circles the additional locations. Follows the observations of hake yield (kg/km) and the predictions obtained by kriging, both on the square root scale. The rightmost plots present the kriging variance

